

QUANTIFICATION OF TOTAL SUSPENDED PARTICULATE (TSP) AND ITS ELEMENTAL CONTENT IN DIFFERENT MICRO ENVIRONMENTS OF A RESIDENTIAL AREA IN LAGOS STATE, NIGERIA

ABSTRACT

The need for comprehensive periodic air quality measurements to assess the extent of airborne particulate matter and trace metals exposure in residential areas in Lagos State in view of the urbanization and different energy cooking source cannot be overemphasized. Total Suspended Particulates (TSP) was collected from July 2016 to April 2017 by gravimetric sampling technique in different indoor-outdoor micro environments of a residential area in Lagos and was analyzed by Atomic Absorption Spectroscopic (AAS) method. High base-line concentrations were obtained with an indoor range of 833.33-1944.45 μgm^{-3} and outdoor range of 1111.11-1944.45 μgm^{-3} during the wet season. During the dry season, it ranged from 1111.11-2777.78 μgm^{-3} in the indoors and 1388.89-2222.22 μgm^{-3} in the outdoors. Elemental concentrations were subjected to enrichment factor analysis (EF) and principal component analysis (PCA) for source identification. EF analysis was used to assess the relative contributions of natural and anthropogenic metals inputs to the air in the area while, PCA identified road dust, combustion activities and marine as the predominant sources of pollutants emission to the environment. To further elucidate the relationship among pollutants in the sampling sites, correlations analysis, cluster analysis and ANOVA were carried out. The results of the correlations analysis and cluster analysis confirmed the results of the EF and PCA. The results of ANOVA showed that, there was no spatial variation in the elemental concentration of most of the metals ($P>0.05$). Conclusively, results obtained showed that, pollutants concentration in most of the sites were higher than safe limits proposed by regulatory limits.

Keywords: Total Suspended Particulate Matter (TSP), Indoor- Outdoor air pollution, cooking fuel and elemental characterization.

1. INTRODUCTION

Atmospheric pollution has generally gained great national and global interest in the recent years due to its large societal burden and various adverse effects (Onabowale and Owoade, 2015). Air pollution can be grouped into two categories: outdoor air pollution (OAP) and indoor air pollution (IAP) (United Nations, 2011). Outdoor air enters and leaves the indoor micro environment by infiltration, natural and mechanical ventilation. When air pollutants from the outdoor air enter, they can either be diluted or accumulated depending on the ventilation condition. The indoor environment can be the home, office, school, workplace or public building. Even time spent within the car or in public transport can be considered as being in an 'indoor' environment (Simoniet *et al.*, 2003). Indoor Air Quality (IAQ) has gained

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great attention in recent years; chiefly due to the ample of time we spend indoors in modern times. People living in urban cities spend 87% indoors and only a mere 6% outdoors (Onabowale and Owoade, 2015). Studies have shown that, indoor sources significantly contribute to personal exposure concentrations experienced in indoor microenvironments. In addition, behaviour of people and ventilation characteristics of homes in different latitudes may significantly affect the concentrations of pollutants in indoor environments and in personal exposures. High temperature and humidity levels can also increase concentration of some pollutants (Ediagbonya *et al.*, 2013).

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Global data shows that, IAP is far more lethal than OAP. It has been estimated that about half a million women and children die each year from indoor air pollution in India (Smith, 2000). In 2004, less than 300 deaths per million people arising from OAP were reported while approximately 2,200 deaths per million people arising from IAP were reported (Omoleet *et al.*, 2014). According to the World Health Organization, an estimated 7 million people were killed by diseases related to indoor and outdoor air pollution alone in 2012. The organization also reported that, 1.6 million people died from cooking stove fumes. Of the 1.6 million deaths, 396,000 deaths occurred in sub-Saharan Africa, with highest incidents occurring in Nigeria (WHO, 2006; Marguliset *et al.*, 2006). Another WHO report posited that 78% of African population used biomass burning to cook and that a third of infant deaths associated with IAP occurred in Africa (WHO, 2007). According to a recent comparative risk study by the World Health Organization (WHO), 28% of the overall disease and deaths is caused by indoor air particulates in developing countries. This may be attributed to the time people in urban areas spend indoors. Apart from death, inhalation of particulate matter in the indoor environment can cause a range of adverse health effects (Onabowale and Owoade, 2015).

The concentrations of total suspended particulate matter in Lagos ambient air have been investigated (Oluyemi and Asubiobjo, 2001). However, there is scanty information on TSP with respect to the different energy cooking source in the indoor micro environment. This study is therefore, necessary considering the potential negative health implications some cooking fuels can pose.

2. MATERIALS AND METHODS

2.1 SITE DESCRIPTION

The study was carried out in Akoka, a residential area in Lagos, Nigeria. The area is surrounded by the Lagos Lagoon and is densely populated with moderate traffic volume and other activities. Prominent in this area are food vendors, business centres, small-scale businesses, artisans, filling stations, banks, tertiary institutions amongst others. Activities in these afore-mentioned are possible sources of pollutants in the environment.

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76 **Figure 1: Map of Akoka Showing the 76 Various Sampling Site**

77 **Table 1: Description of Sampling Sites Assessed**

78 AKO.CI and AKO.CO = Akoka charcoal indoor and outdoor AKO.KI and AKO.KO = Akoka
79 kerosene indoor and outdoor AKO.GI and AKO.GO = Akoka gas indoor and outdoor
80 AKO.VO = Akoka vehicular outdoor.

81 2.2 SAMPLE COLLECTION

82 Sample was systematically collected directly by the use of a gravimetric high volume air
83 sampler. The particulates were collected from the ambient air on the pre weighed whatmann
84 cellulose filter paper using a portable high volume gravimetric sampler (Hi-Q CF – 901).
85 2.5L/min volume air was pumped through the filter paper for eight hours on each occasion.
86 The sampling height in the course of this work was 1.6 m. The sampler was closely
87 monitored throughout the duration of sampling to ensure accurate sampling collection timing,
88 prevent battery failure and vandalism. For each sampling, the filter and cassette was humidity
89 conditioned (equilibrated) in a charged desiccator for 24 hours and weighed before and after
90 sampling. After sampling, the loaded filters were stored in sealed polythene bags and taken to
91 laboratory for sample preparation and elemental analysis by Atomic Absorption Spectroscopy
92 (AAS).

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93 2.3 SAMPLE PREPARATION

94 The loaded filter paper was carefully placed inside a 100ml beaker. 10ml of HNO_3 was added
95 and heated at 150°C in a fume cupboard. The sample was intermittently spiked with 5ml
96 perchloric acid after an hour and was heated for three hours until a clear solution was
97 observed. The digest was cooled, filtered into 50 ml standard volumetric flask and diluted
98 with distilled water to the ml mark in the 50 ml volumetric flask. Elemental analysis was then
99 carried out on the final digest using Perkin Elmer A Analyst 400 atomic absorption
100 spectrophotometer.

101 3. RESULTS AND DISCUSSION

102 High concentrations of TSP were obtained with an indoor range of $833.33\text{--}1944.45\mu\text{g m}^{-3}$ and
103 outdoor range of $1111.11\text{--}1944.45\mu\text{g m}^{-3}$ during the wet season. During the dry season, it
104 ranged from $1111.11\text{--}2777.78$ in the indoors and $1388.89\text{--}2222.22\mu\text{g m}^{-3}$ in the
105 outdoors. These values are about 3 to 11 fold the $250\mu\text{g m}^{-3}$ stipulated by the Federal Ministry
106 of Environment (FMNEV) and World Health Organization (WHO) statutory limits of
107 $250\mu\text{g m}^{-3}$ and 150 to $230\mu\text{g m}^{-3}$ respectively (WHO, 2005) thus, they clearly violate the
108 statutory limits. The high level of TSP in the air is probably from both natural and
109 anthropogenic sources with the latter being more dominant.

110 **Table 2: Comparison of Indoor-Outdoor Total Suspended Particulate Matter Results**
111 **($\mu\text{g m}^{-3}$) of this Study with Others**

112

113 The high levels of pollutants measured in this study and other studies (Table 2) is a signature
114 of anthropogenic activities and is an indication that, the indoor-outdoor micro environments
115 of many urban and rural areas are highly polluted and therefore calls for urgent environment
116 pollution control.

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117 3.1 ELEMENTAL CONCENTRATION AND ENRICHMENT FACTOR

118 When there is high concentration of heavy metals in airborne particulate matter, it becomes a
119 thing of serious environmental concern because of the implications it poses to man and its
120 environment. It therefore becomes necessary to monitor their concentration through air
121 quality studies.
122 The mean elemental concentrations together with the mean enrichment factors calculated for
123 the indoor-outdoor micro environments during the wet and dry seasons of the sampling period
124 are presented in Table 3-6.

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131 **Table 3: Mean Elemental Concentration and Enrichment Factor of Indoor Total**
132 **Suspended Particulate Matter during the Wet Season**

133

134 **Table 4: Mean Elemental Concentration and Enrichment Factor of Outdoor Total**
135 **Suspended Particulate Matter during the Wet Season**

136

137 **Table 5: Mean Elemental Concentration and Enrichment Factor of Indoor Total**
138 **Suspended Particulate Matter during the Dry Season**

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141 **Table 6:Mean Elemental Concentration and Enrichment Factor of Outdoor Total**
142 **Suspended Particulate Matter during the Dry Season**

143
144 **3.1.1 Elemental Concentration**

145 Generally, it was observed that, the concentrations of the metals analyzed were higher
146 indoors than outdoors. The increased indoor pollutants are in agreement with the studies of
147 Environmental Protection Agency. EPA studies have shown that levels of air pollutants
148 indoors may be anywhere from 2 to 5 times greater than outdoors and in some cases, more
149 than 100 times greater (Pau, 2008).Also, according to Chen and Zhao (2011), in combined
150 indoor and outdoor air quality studies, more than 2/3 authors found indoor air pollutant
151 concentration higher than outdoor pollutant concentrations. This is probably due to a higher
152 dispersion rate in the outdoor micro environment compared to the corresponding indoor
153 micro environment.

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154 It was also observed that, the concentrations of the metals measured during the wet season
155 were relatively lower than that measured during the dry season in the various sites and this
156 might be attributed to metrological factors such as: (i) lower air temperature (ii) higher
157 humidity, (iii) higher wind speed and (iv) increase in cloud covers and consequently, increase
158 in the number of rainfall days experienced during the wet season. A lower temperature and a
159 higher humidity cause trapping of particulates in water molecules in the atmosphere (Okuoet
160 al., 2017). These trapped particulates are later washed down via rainfall or droplets. Also, the
161 greater the wind speed, the greater the turbulence and the more rapid and complete the
162 dispersion of contaminants in the air (Guttikunda and Gurjar, 2012).

163 A similar study by Nasir et al., (2015) reported indoor-outdoor relationships of trace metals in
164 PM in an urban and two rural areas of Pakistan. The concentrations of most of the metals in
165 the indoor micro environment were higher than the outdoor micro environment in their study.
166 This trend was also observed in this study. The reason for the observed lower concentrations
167 in the outdoor micro environments of both studies is probably due to a higher dispersion rate.

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168 **3.1.2 Enrichment Factor (EF)**

169 In order to discriminate the natural sources from the anthropogenic sources of atmospheric
170 metals emission, enrichment factors (EFs) were calculated as:

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173 Where E is the element of interest and R, the reference material

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178 The EF for Na a marker element for marine, was >10 in all the sites in both seasons. This
179 implies severe enrichment in the various sites. The high EF of Na should be due to the
180 pronounced effect of sea breeze in the region as Lagos is situated on the coast of the Atlantic
181 Ocean. The EF of Ca and K was less than 3 in virtually all the sites. This is an indication that
182 the main source of emission of these metals is natural. The EF of Cu, Pb, Cd and Ni were \geq
183 25 in all the sites in the two seasons. The high enrichment of these metals suggests that, their
184 dominant source.

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s are anthropogenic and a variety of pollution emissions such as vehicular exhaust/oil combustion, waste burning, smoking, paints, amongst others, might have contributed to their loading in the ambient air.

3.2 MULTIVARIATE ANALYSIS

The mean elemental concentrations of TSP measured in the various sampling sites were subjected to analysis of variance (ANOVA), inter-metallic correlation, principal component analysis (PCA) and cluster analysis using SPSS statistical package. This was done to elucidate relationships among the sampling sites and to identify the sources contributing to the metals level in the area.

3.2.1 Analysis of Variance (ANOVA)

It was used to compare the mean concentrations of the metals in the various sampling sites so as to determine whether, they were significantly different or not. Table 7 and 8 shows the spatial variations of the metals in the indoor and outdoor microenvironments for both seasons.

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201 **Table 7: Analysis of Variance (ANOVA) for the Wet Season**

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203 **Table 8: Analysis of Variance (ANOVA) for Dry Season**

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205 The results of the ANOVA as shown in Table 7 and 8, show that apart from Pb, Cd and K that
206 showed significant variation in the indoor micro environment and Pb, Cd, Cu and K in the
207 outdoor micro environment ($p < 0.05$), there were no significant variation of the metals
208 distribution in the various sampling sites. During the dry season, apart from Pb, Cd and K that

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209 howed significant variation in both the indoor and outdoor micro environment ($p < 0.05$), it was
210 observed that variation of the metals distribution in the various sampling sites were not
211 significant. Factors responsible for spatial and ambient concentrations of air pollutants which
212 could also be responsible for the trend in this area include: emission strength, emission rate,
213 emission conditions and atmospheric dispersion conditions (Ukbebo *et al.*, 2012).

214

215 3.2.2 Seasonal Variations of Elemental Concentration

216 In order to determine the seasonal variation of the elemental concentrations in the various
217 sampling sites, the concentrations of the analyzed metals for the two seasons were subjected to
218 ANOVA and are presented in Table 9 and 10.

219 Table 9: Seasonal Variations of Elemental Concentration in the Indoors

220

221 Apart from Pb that showed significant variation in site AKO.KI ($p < 0.05$), there was no
222 significant variations in the concentrations of the metals in the indoor micro environment
223 sampling sites during the wet and dry seasons of the sampling period. The seasonal significant
224 variation observed in Pb concentration in site AKO.KI is probably an indication that,
225 anthropogenic source(s) of emission during the dry season is more prevalent than the wet season.

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226 As mentioned earlier, metrological factors is yet another reason for the observed seasonal
227 variations.

228

229 **Table 10: Seasonal Variations of Elemental Concentrations in the Outdoors**

230

231 In the outdoor micro environment of site AKO.CO, apart from K that is significantly different
232 ($p < 0.05$), there was no seasonal variations ($p > 0.05$) in the analyzed metals. Metrological factors
233 and a higher rate of emission during the dry season, is probably a reason for the observed
234 increase in the concentration of K in the season. In site AKO.KO there was no significant
235 variations in the concentrations of the metals except in Cd and Ca. The reason for the variation

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236 observed in these two metals is as mentioned in the case of K. In site AKO.GO, seasonal
237 variation was observed only in Pb and K. Again, the reason is as stated above.
238 There was no significant difference in the concentrations of most of the metals analyzed for the
239 two seasons. This seasonal pattern is similar to the seasonal pattern observed for metals in TSP at
240 Guangzhou and Foshan (Xiao *et al.*, 2015).

241

242 3.2.3 Correlation Analysis

243 The possible sources around the sampling sites were qualitatively identified from the correlation
244 matrix by analyzing the correlation between elements. Correlations among metals indicate a
245 common source of origin as shown in Table 11 and 12.

246

247 **Table 11: Inter-metallic Correlations of Indoor-Outdoor Samples**

** . Correlation is significant at the 0.01 level (2-tailed).

* . Correlation is significant at the 0.05 level (2-tailed).

248 The results of inter-elemental correlation matrix and during the wet season showed that, there
249 was strong correlation between Pb-Cd (-0.928), Pb-K (0.690), Cd-Ca (0.759) and Fe-Na (0.712)
250 in the indoors, Pb-Cd (-0.839), Pb-Cu (0.650), Pb-Fe (0.652), Fe-Ni (0.808) and Fe-Na (0.614) in
251 the outdoors. The strong correlation of these metals suggests they may have a common origin.
252 The correlations between Pb, Cd, K and Ca suggests road dust source (i.e., sum of soil dust and
253 automobile source type). While that between Fe and Na may be attributed to crustal dust. In the
254 outdoor samples, there was a strong correlation between Pb-Cd (-0.839), Pb-Cu (0.650), Pb-Fe
255 (0.652), Fe-Ni (0.808) and Fe-Na (0.614). The strong correlation between Pb, Cd, Cu, Ni and Fe

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256 strongly suggests vehicular emission as a prominent anthropogenic source of emission in this
257 area. Emissions from vehicle exhaust and wear abrasion are both important sources of these
258 metals. Again, previous studies have pointed out that, in addition to road dust, wear debris from
259 brake linings and tyres as well as diesel engine emissions are main sources of Fe in areas near
260 traffic emissions (Amato *et al.*, 2009). This therefore confirms the traffic volume in the area.

261

262 **Table 12 Inter-metallic Correlations of Indoor-Outdoor Samples**

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263 **Correlation is significant at the 0.05 level (2-tailed).

264 *Correlation is significant at the 0.05 level (2-tailed).

265 During the dry season, the result of the inter-metallic matrix showed that in the indoors, there
266 was strong correlation between Na-Cu (0.783), Ni-Ca (0.673) and (0.800). These relationships
267 suggest road dust. While in the outdoors, a strong correlations between Pb-Cd (-0.655), Pb-Cu
268 (0.719), Pb- Ni (0.585), Pb-Na (0.683), Cu-Ni (0.698), Cu-Fe (0.586), Ni-Fe (0.626), Ni-Na
269 (0.776), Fe-Na (0.805), Cd-Ca (0.596) and Cu-Na (0.756) was observed. The correlation
270 between Pb-Cd, Pb-Cu, Pb-Ni, Cu-Ni, Cu-Fe suggests vehicular source emission type since these
271 metals are mainly emitted either from fuel, tyre wear/tear or corrosion of car steel parts. While
272 the correlation between Ni-Fe, Ni-Na and Cd-Ca suggests road dust source (i.e., sum of soil dust
273 and automobile source type) as a major source of emission in the sites. The correlation between
274 Fe and Na may be attributed to crustal dust.

275 3.2.4 Principal Component Analysis

276 In order to identify the source(s) contributing to the metals emissions at the sampling sites, the
277 obtained data were further subjected to PCA. To determine the number of factors to retain in the
278 results, the values of variance after rotation were examined and only factors with eigenvalues ≥ 1
279 were considered significant as shown in Table13-16.

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280 **Table 13: Rotated Factor Loading for TSP in the Indoor Sites During the Wet season**

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281

282 In the indoor micro environment, two major factors were identified. Factor 1 (PC1) loads heavily
283 on Cu, Ni, Fe, Ca, Na and K. This source may be attributed to the combination of marine and

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284 road dust. Factor 2 (PC2) is loaded primarily by Pb, Cd, Ca, Cu and K. This is likely from
285 anthropogenic activities and may be attributed to biomass/waste burning and construction.

286

287 **Figure 2: Indoor Samples Component Plot in Rotated Space during the Wet Season**

288

289 **Table 14: Rotated Factor Loading For TSP in the Outdoor Sites during the Wet Season**

290

291 Factor 1 (PC1) loads heavily on Pb, Cd and Cu. This is may be attributed to waste burning.
292 Factor 2 (PC2) is loaded primarily by Ni, Fe, Ca and K and may be related to construction
293 activities and biomass burning. Factor 3 (PC3) loads heavily on Cu, Fe and Na. This factor
294 represents a combination of vehicular and marine. Cu and Fe is marker element for
295 brake/tyre/car parts wear and tear and thus can serve as indicators of traffic re-suspension
296 (Amato *et al.*, 2009).

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298 **Figure 3: Outdoor Samples Component Plot in Rotated Space during the Wet Season**

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309 **Table 15: Rotated Factor Loading For TSP in the Indoor Sites during the Wet season**

310

311 Factor 1 (PC1) loads heavily on Cd, Cu, Ni and Na. This factor may be attributed to the
312 combination of waste burning and marine. Factor 2 (PC2) is loaded primarily by Cd, Fe and K.
313 This is most likely related to road dust and biomass burning. Factor 3 (PC3) loads heavily on Pb,
314 Ni and Ca. This factor may be attributed to vehicular and construction activities.

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316 **Figure 4:Indoor Samples Component Plot in Rotated Space during the Dry**

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326 **Table 16: Results of the Rotated Factor Loading for TSP in the Outdoor Sites during the**

327 **Dry Season**

328

329 Factor 1 (PC1) loads heavily on Pb, Cu, Ni, Fe and Na. This may be attributed to road dust and
330 sea spray (marine). Factor 2 (PC2) is loaded primarily by Pb, Cd and Ca. This may be related to
331 waste burning and construction activities. Factor 3 (PC3) loads heavily on K. This factor is
332 probably related to biomass burning.

333

334 **Figure 5: Outdoor Samples Component Plot in Rotated Space during the Dry Season**

335

336 3.2.5 Cluster Analysis

337 The result of the cluster analysis carried out with the aid of cluster package using Euclidean
338 distance and complete linkage farthest neighbors as a measure of correlation was able to confirm
339 the sources as identified by the PCA analysis.

340 During the wet season, the result showed significant clustering of Cd, Cu, Pb and Ni; Fe, Ca and
341 K and Na in the indoor micro environment. While a significant clustering of Pb Cd and Cu; Ni,
342 Fe, Ca and K and Na was observed in the outdoor micro environment. Cluster 1, 2 and 3 may be
343 relevant to vehicles or waste combustion, crustal dust and marine respectively in both indoor and
344 outdoor micro environment. During the dry season, the indoor result showed significant
345 clustering of Cd, Cu, Pb, and Ni; Ca and Na; Fe and K. While; a significant clustering Pb, Ni, Cd
346 and Cu; Fe and Ca; Na and k was observed. Cluster 1, 2 and 3 may be relevant to vehicles or

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347 waste combustion and crustal dust in the indoors. And in the outdoors, cluster 1, 2 and 3 may be
348 relevant to vehicles or waste combustion, crustal dust and marine.

349 CONCLUSION

350 The results of this study identified road dust, marine and combustion activities as the major
351 sources of PM and metals in this area. It also revealed that, the concentration of these pollutants
352 in most of the sites were higher than safe limits proposed by regulatory bodies and this can have
353 significant implications on the vulnerable groups (elderly, children, pregnant women) and public
354 health in general. Of the three cooking fuels in this study, charcoal generated the highest TSP
355 while, gas stove generated the least. The use of clean cooking fuel in homes should therefore be
356 adopted where possible.

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357 REFERENCES