Variance in Protease, Dehydrogenase, Phosphatase and Respiratory Activities during Phytoremediation of Crude Oil Polluted Agricultural Soil Using Schwenkia americana and Spermacoce ocymoides
ABSTRACT
 Aims: To investigate the variation in the activities of some soil enzymes and microb respiration during phytoremediation of crude oil polluted agricultural soil. Study design: Indigenous plants of a crude oil polluted agricultural farmland we harvested and identified. Two species (<i>Schwenkia americana</i> and <i>Spermacoce ocymoide</i> were selected for the study. Nursery was set up using sterile soil and mature and viat seeds of the selected species, and germinated seedlings were transplanted into an 8 potted homogenized polluted soil for remediation. Place and Duration of Study: Polluted agricultural soil from Ogoniland Nigeria, Univers of Port Harcourt ecological garden, between May 2017 and February 2018. Methodology: Protease activity was determined based on the amino acids released affincubation of the soil with sodium caseinate. Phosphatase activities determination w based on determining the extent of degradation of <i>p</i>-nitrophenol phosphate (PNPP) by the samples. Dehydrogenase activity was determined based on the estimation of the rate reduction of 2,3,5- triphenyltetrazolium chloride (TTC) to triphenyl formazan (TPF) in so after incubation. The substrate induced method was adapted to estimate the respirator activity. Organic matter was determined by loss of weight on ignition method. Results: Dehydrogenase activities of the remediated groups increased after 4 weeks the decreased at the end of the remediated groups reduced over time while the soil microbar respiratory activity reduced at the end of the 12 weeks remediation. A trend between organ matter and enzyme and respiratory activities can reflect soil quality, and soil enzyme activities can directly reflect the metabolic need and nutrient availability of soil microorganisms. The substrate induced and nutrient availability of soil microorganisms. The substrate induced and nutrient availability of soil microorganisms. The substrate herease and nutrient availability of soil microorganisms. There

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1. INTRODUCTION

Soil is an abode for life [1] and its contamination by petroleum hydrocarbons emanating from commercial exploration and spillage [2] from the activities of the petroleum industry (including oil exploration, drilling, production, transportation, processing and storage), well blow-outs, pipeline rupture, tanker accidents, and pipeline vandalization by saboteurs and hoodlums poses a risk challenge in many oil producing areas ascribable to their environmental consequences to man [3].

Oil pollution dreadfully affects the soil ecosystem through adsorption and surface assimilation of soil particles purveying of an excess carbon which might be unattainable for microbial use and the

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investiture of a constraint in soil nutrients [4]. During oil spillages, non-organic compounds, 26 27 carcinogens, and growth inhibiting chemicals obtainable in crude oil are introduced to the environment 28 [3], and protracted exposure to acute oil contamination could result to the instigation of kidney and 29 liver diseases, mutilation of bone marrow and intensified risk of cancer [5]. There is a proportional 30 reduction in contaminant extraction and biodegradation as the interaction between particles of soil and pollutants increase [6]. Biodegradation makes use of bacteria, fungi or various biological means 31 32 to disintegrate materials. Microorganisms possess a great ability to metabolize degradable 33 contaminants by employing them as energy source and/or converting them to non-toxic product such 34 as carbon dioxide, biomass and water. This relies on the nature and amount of hydrocarbons present 35 [7].

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37 Microbial and enzymatic activities of the soil can reveal succinctly the quality of soil [8]. The activities 38 of soil enzymes can be used to reveal the metabolic need and nutrient availability of soil 39 microorganisms which are essential in the processing and recovery of key nutrients from detrital 40 inputs and accumulated soil organic matter [9]. Extracellular enzymes such as proteases, 41 dehydrogenases and phosphatases are involved in the process of organic matter decomposition and 42 cycling of key elements such as carbon, nitrogen and phosphorus [10]. Studies have revealed that 43 enzyme activities in the soil are related to heavy metal contamination. Almost all enzyme activities in 44 soils are significantly reduced by 10 to 50 times with the increase of the concentration of heavy metals 45 in the soil [11]. Heavy metal toxicity affects microbial population size, diversity, and activity and also 46 affects their genetic structure. It also alters the nucleic acid structure, disrupts cell membrane, and causes functional disturbance thereby inhibiting the enzyme activity and oxidative phosphorylation. 47 48 causing lipid peroxidation and altering osmotic balance and protein denaturation [12]. This study thus 49 assays for the presence of some soil enzymes in crude oil polluted agricultural soil and their activities 50 with respect to remediation of the soil using Schwenkia americana and Spermacoce ocymoides.

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52 2. METHODOLOGY

54 A polluted agricultural farmland located in Ogoniland, Nigeria was identified in Bodo community, 55 Gokana L.G.A. of Rivers state and assessed to ascertain the types of contaminants involved and to 56 determine the most appropriate technologies for its restoration. In the assessment, the site was 57 mapped to determine its physical characteristics, size and location of contaminants as well as the 58 plant ecological community. Thereafter, indigenous plants of the polluted site were harvested and 59 taken to the Department of Plant Science and Biotechnology, University of Port Harcourt, Nigeria for 60 identification. Two species (Schwenkia americana and Spermacoce ocymoides) were selected for the 61 study owing to existing reports on their survival ability in polluted environments. Soil samples from 62 crude oil polluted site and agricultural soil from natural matrix within the University of Port Harcourt 63 were collected following the described method [13]. Nursery was set up using sterile soil and mature 64 and viable seeds of the selected species. Three to four weeks after germination, 4 seedlings each of 65 the plants were transplanted into an 8 kg potted homogenized polluted soil set up in triplicate 66 alongside unvegetated polluted and unpolluted control soils. Soil sampling was carried out prior to the transplant and subsequently at 4th, 8th and 12th week. Fresh soil samples were collected and taken 67 68 immediately to the laboratory. The activities of acid and alkaline phosphatase, dehydrogenase and 69 protease were assayed and the organic matter content of the soil was determined.

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The method [14] with modification was employed for all enzyme assays and respiratory activity
 performed in triplicate and compared to controls while soil organic matter was determined by loss of
 weight on ignition method [13].

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75 The assay for protease activity is based on determining the amino acids released after incubation of 76 the soil with sodium caseinate for 2 hours at 50°C using Folin-Ciocalteu reagent. Two grams of moist, 77 sieved (2 mm) soil was weighed into a 15-mL centrifuge tubes designated as test and control. Aliquot 78 (5 mL) of 1% substrate, prepared a night before and kept in a refrigerator, was added to the test 79 tubes. For the controls, only 5 mL of TRIS HCI buffer at pH 8.1 was added. The tubes were shaken for 80 2 hours at 50°C and cooled immediately in cold water. An aliquot of 2 mL 17.5% trichloroacectic acid 81 was added into test and control tubes and centrifuged at 3000 rpm for 2 minutes. The supernatant (2 82 mL) was dispensed into test tubes, and 3 mL 1.4M NaSO₄ was added in both the test and control tubes. The tubes were shaken thoroughly and 1 mL of dilute Folin-Ciocalteu reagent, prepared by 83 84 diluting three times, was added and the content of the tubes centrifuged at 200 rpm for 2 minutes. An

85 aliquot from each tube was taken and read on a spectrophotometer at a wavelength of 578 nm. TRIS 86 HCl buffer at pH 8.1 was used as blank and calibration curves were prepared.

Enzyme activity (mg tyrosine kg⁻¹ dry matter h⁻¹) = (C x 10.5) / dw

where C = measured tyrosine concentration; dw = dry weight of 2g moist soil. 88

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90 Acid and alkaline phosphatase activities were assayed based on the determination of the extent of degradation of p-nitrophenol phosphate (PNPP) by the samples. One gram each of air-dried and 91 92 homogenized soil samples was placed in centrifuge tubes and 0.25 mL toluene was added, and 93 placed in a fume chamber for 10 minutes. Thereafter, 4 mL of ACP (or ALP) buffer solution was added followed by the addition of 1 mL of ACP (or ALP) PNP substrate. The contents of the tubes 94 95 were incubated for 1 hour at 37°C. Aliquots of 1 mL of 0.5 M CaCl₂ and 4 mL of 0.5 M NaOH were added to the tubes, whose contents were shaken with a rotator mixer for 3 minutes and centrifuged at 96 97 4000 rpm for 10 minutes. The supernatant was read spectrophotometrically at 485 nm using ACP (or 98 ALP) buffer solution as blank. The controls were prepared similarly but without the substrate while the 99 calibration curve was prepared using standard p-nitrophenol solution. The p-nitrophenol per mL of 100 filtrate was calculated by reference to the calibration curve. The ρ -nitrophenol released after 101 incubation was used to calculate the enzyme activity as follows:

Enzyme activity (mmol PNP kg⁻¹ dw h⁻¹) = $(C \times V) / (dw \times SW \times t)$ 102 where C = measured concentration of ρ -nitrophenol in μ g ml⁻¹ filtrate; V = total volume of soil 103 104 suspension in ml; dw = dry weight of 1 g moist soil; SW = weight of the soil sample used; and t = 105 incubation time in hours.

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107 The assay of dehydrogenase (DH) activity is based on the estimation of the rate of reduction of 2,3,5-108 triphenyltetrazolium chloride (TTC) to triphenyl formazan (TPF) in soils after incubation at 30 °C for 24 109 hours. Five grams of field-moist soil was prepared in centrifuge tubes designated as test and control 110 tubes. Using Eppendorf pipette, 5 mL of the 1% TTC was added to the test tubes while 5 mL of TRIS 111 HCl buffer at pH 7.4 was added to the control tubes. The content of the tubes was mixed thoroughly 112 and incubated at 30 °C for 3 d. After incubation, 20 mL of methanol was added in all the tubes and 113 shaken for 5 minutes on a turnover shaker at room temperature and then centrifuged at 300 rpm for 3 114 minutes. An aliquot from each of the tubes was taken and analyzed using a spectrophotometer at a 115 wavelength of 485 nm, using TRIS HCl at pH 7.4 as blank. The ρ -nitrophenol contents of the filtrates 116 were extrapolated from the calibration curve. The amount of TPF formed was used to evaluate the 117 enzyme activity as follows:

Enzyme activity (mg TPF kg⁻¹ d⁻¹) = (C x v) / (dw x SW x t) where C = measured concentration of TPF in mg ml⁻¹ filtrate; v = total volume of soil suspension in 119 120 ml; dw = dry weight of 1 g moist soil; SW = weight of the soil sample used; and t = incubation time in 121 hours.

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123 The substrate (glucose) induced method was adapted to estimate the respiratory activity. A 10 g 124 screened (2 mm sieve) soil sample was added into the outer jar of respiratory flask. An aliquot of 2.5 125 mL 0.2 M NaOH was added into the inner jar. Glucose solution (1 mL) was added to the soil. For the 126 control flasks, screened (2 mm) 10 g of soil sterilized in an autoclave at 121 °C for 20 minutes was 127 used. An aliquot of 2.5 mL 0.2 M NaOH was added with 1 mL 40% glucose as describe above. The 128 flasks were allowed to stand for 24 hours at 25 °C. Afterwards, all the NaOH was transferred to a 25 129 mL beaker containing 1 mL 1 M BaCl₂ and the solution was titrated using 0.1 M HCl and 130 phenolphthalein indicator. The titre values were noted for calculation.

131 To calculate the mass of CO2 generated:

132 $CO_2.C (mg) = [(HCI_b - HCI_s) / 1000 mL/L] \times HCI molarity (mol/I) \times 12g C/mol \times 1000 mg/g$ 133 Where HCl_b = mI HCl used in titration of blank; HCl_s = mI HCl used in titration of sample; $CO_2 \cdot C =$ 134 mass of CO₂-carbon generated (mg).

135 This simplifies to: 136

$$CO_2.C (mg) = (HCI_b - HCI_s) \times 12$$

137 138 To determine the organic matter content of the soil samples, five grams of sieved (2 mm) soil samples 139 were weighed into crucibles. The crucibles with the soil samples were placed in a drying oven, set at 140 105 °C and allowed to dry. After 4 hours, the crucibles were removed from the drying oven and placed 141 in a dry atmosphere. When cooled, the crucibles with the soil samples were weighed to the nearest

- 142 0.01 g. Again, the crucibles with the dried soil samples were placed in a muffle furnace, set at 400 °C.
- 143 After 4 hours of ashing, they were removed from the muffle furnace, cooled in a dry atmosphere, and
- reweighed to the nearest 0.01 g.
- 145 The percentage organic matter is given by: 146 % OI

$$\% OM = [(W_1 - W_2) / W_1] \times 100$$

147 where W_1 = the weight of soil at 105°C; W_2 = the weight of soil at 400°C.

148 3. RESULTS AND DISCUSSION

149 The protease activities of the various soil samples are presented in Table 1. Compared to baseline 150 values, the protease activities of the remediated groups reduced over time. This may be due to the 151 inhibitory influence of the remediating plants on the soil microorganisms. It may however be due to 152 the limiting effect of nutrients in the pots, since they have been depleted over time, with the resultant 153 reduction in microbial activity. The later argument may account for the reduction observed for the 154 unpolluted group. The former contention can be substantiated by the findings of [15] that plant 155 extracts of *M. alternifolius* and other plants inhibited the growth of certain fungi and bacteria, with *M.* 156 strongly inhibiting the fungi P. chrysogenum and bacteria Escherichia coli, alternifolius 157 Staphylococcus aureus and Salmonella typhi. This study also revealed a trend between organic 158 matter and enzyme and respiratory activities. Pearson's correlation coefficient (PCC, a measure of the 159 linear correlation or dependence between two variables) of -1.00, -0.98 and -0.80 (Table 7) for 160 protease activity in unpolluted control, polluted control and S. americana treated groups, respectively, 161 showed substantial negative correlations with organic matter (OM) where as the soil treated with S. 162 ocymoides showed fair positive correlation (+0.47 PCC) as shown in Table 7.

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164Table 1.Protease activity (in mg tyrosine kg^{-1} dry matter h^{-1}) of unpolluted control,165polluted control, Schwenkia americana and Spermacoce ocymoides.

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Group	Before	Week 4	Week 8	Week 12
Unpolluted	32.9021±3.86 ^a	38.6255±0.77 ^{a*}	20.3461±1.59 ^{a*}	11.9031±1.00 ^{a*}
Control				
Polluted	44.4372±0.77 ^b	21.9986±8.06 ^{b[*]}	13.6596±43.47 ^{a,b*}	2.6944±2.88 ^{b*}
Control				
Schwenkia	44.4372±0.77 ^b	24.1303±3.32 ^{b*}	10.5389±1.18 ^{b*}	1.3813±0.47 ^{b*}
americana				
Spermacoce	44.4372±0.77 ^b	23.6187±16.81 ^{a,b*}	6.2756±5.56 ^{b*}	2.4046±1.19 ^{b*}
ocymoides				

167 Values are mean ± standard deviations of triplicate determinations.

168 Values in the same column with different letters (a,b) are significantly different at P = .05.

169 *P = .05 compared to the corresponding values before treatment.

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171 The dehydrogenase activities of the various soil samples are presented in Table 2. Compared to 172 baseline values, the dehydrogenase activities of the remediated groups showed a significant (p < 0.05) 173 rise in activity after 4 weeks but reduced at the end of remediation. The increase may have been a 174 result of an initial increase in microbial population within the first 4 weeks which afterwards reduced 175 with depletion of carbon source or available nutrients, since they have been depleted over time, with 176 the resultant reduction in microbial activity. Though there might be available nutrients in the 177 unpolluted group, the absence of carbon source may account for the insignificant activities observed. 178 [16] Reported an undesirable reduction in the dehydrogenase activity and associated that with the low 179 activities of microorganisms in polluted soil. [17] Made a clearer and more acceptable report that both 180 the microbial population, activity of the microbial population and the kind of microbe present in the soil 181 determine the enzyme activity. This trend as observed in Table 2 follows similar trends [18], [19], [20]. 182 Substantial positive correlation (+0.96 PCC) for dehydrogenase activity and OM was observed only in 183 the soil treated with S. americana. While the unpolluted control and soil treated with S. ocymoides 184 showed a fair positive correlation of +0.36 and +0.55 respectively, the polluted control soil showed a 185 substantial negatively correlation (-0.90 PCC).

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191Table 2.Dehydrogenase activity (in mg TPF kg⁻¹ d⁻¹) of unpolluted control, polluted192control, Schwenkia americana and Spermacoce ocymoides.

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Group	Before	Week 4	Week 12
Unpolluted Control	1.1837±0.95 ^a	7.6737±6.72 ^a	0.6091±0.08 ^a
Polluted Control	0.1528±0.21 ^b	11.4075±6.66 ^{a*}	0.6318±0.11 ^ª
Schwenkia americana	0.1528±0.21 ^b	5.2486±2.72 ^{a*}	0.9236±0.40 ^b
Spermacoce ocymoides	0.1528±0.21 [⊳]	7.4165±0.61 ^{a*}	0.1450±0.06 ^c

194 Values are mean ± standard deviations of triplicate determinations.

195 Values in the same column with different letters (a,b) are significantly different at P = .05.

196 **P* = .05 compared to the corresponding values before treatment.

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198 The acid and alkaline phosphatase activities of the various soil samples are presented in Tables 3 199 and 4. Compared to baseline values, the acid phosphatase activities (Table 3) of the remediated 200 groups reduced over time. This may be due to the inhibitory influence of the remediating plants on the 201 soil microorganisms. It may however be due to the limiting effect of nutrients in the pots, since they 202 have been depleted over time, with the resultant reduction in microbial activity. The later argument 203 may account for the reduction observed for the polluted and unpolluted groups, since the pots were 204 unvegetated. However, the increase observed at week 8 for unpolluted group may indicate a rise in 205 peak in microbial activity which may have reduced owing to the depletion in available nutrients. 206 Likewise, compared to the baseline values, alkaline phosphatase activities of the remediated groups, 207 as shown in Table 4, reduced over time albeit a recorded increase in S. ocymoites treated group at 208 week 8. The reduction may be due to the hampering influence of the remediating plants on the soil 209 microorganisms. Nonetheless, it may be due to the limiting effects of nutrients in the pots as depletion 210 may have taken place over the period of time, thus resulting to reduction in microbial activity. If the 211 later argument is true, it may therefore account for the reduction observed for unpolluted and polluted 212 groups. However, the population and/or the presence of certain microorganisms specific for alkaline 213 phosphatase secretion may have influenced the increase in activity recorded in S. ocymoides treated 214 group at week 8. This finding is supported by the report [17], that microbial population, activity and the 215 kind of microbe present in the soil determine the enzyme activity. Acid phosphatase activity showed a 216 substantial positive correlation with OM for the unpolluted control and S. americana treated soil (+0.84 217 and +0.80 PCC, respectively). However, whilst the soil treated with S. ocymoides showed a fair 218 positive correlation (+0.55PCC); the polluted control soil indicated almost no correlation (+0.05 PCC). 219 On the other hand, alkaline phosphatase activity revealed a fair positive correlation with OM for 220 unpolluted control, polluted control and soil treated with S. americana (+0.61, +0.52, and +0.36 PCC, 221 respectively), and its correlation with OM for S. ocymoides treated soil revealed a fair negative 222 correlation of -0.57.

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Table 3: Acid phosphatase activity (mmol PNP kg⁻¹ dw h⁻¹) of unpolluted control, polluted control, *Schwenkia americana* and *Spermacoce ocymoides*.

Group	Before	Week 4	Week 8	Week 12
Unpolluted	1.5190±1.06 ^a	1.6357±0.15 ^a	7.5831±1.22 ^a *	0.0485±0.06 ^a
Control				
Polluted	5.4736±1.74 ^b	1.5739±1.00 ^a	2.8370±4.61 ^{a,b,c}	0.1502±0.04 ^{a*}
Control				
Schwenkia	5.4736±1.74 ^b	4.5386±1.21 ^b	6.1315±1.72 ^c	0.2076±0.39 ^{b,c*}
americana				
Spermacoce	5.4736±1.74 ^b	6.5304±1.51 [°]	0.1773±0.50 ^b *	0.2127±0.25 ^{a,c*}
ocvmoides				

227 Values are mean ± standard deviations of triplicate determinations.

228 Values in the same column with different letters (a,b) are significantly different at P = .05.

229 **P* = .05 compared to the corresponding values before treatment.

Table 4: Alkaline phosphatase activity (mmol PNP kg⁻1 dw h⁻1) of unpolluted control, polluted control, *Schwenkia americana* and *Spermacoce ocymoides*.

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Group	Before	Week 4	Week 8	Week 12
Unpolluted Control	3.2944±1.11 ^a	0.1846±0.16 ^{a,b*}	3.2186±1.20 ^a	0.6330±0.43 ^{a*}
Polluted	4.3601±1.06 ^a	0.5912±0.26 ^{a,c*}	3.3636±1.05 ^ª	1.2072±0.96 ^{a*}
Schwenkia	4.3601±1.06 ^ª	0.4942±0.07 ^{c*}	0.1835±0.04 ^{b*}	0.6294±0.31 ^{a*}
Spermacoce ocymoides	4.3601±1.06 ^a	0.2753±0.64 ^{b*}	11.4072±2.44 ^{c*}	0.8385±0.69 ^{a*}

234 Values are mean ± standard deviations of triplicate determinations.

235 Values in the same column with different letters (a,b) are significantly different at P = .05.

236 *P = .05 compared to the corresponding values before treatment.

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238 Assessment of oxidation of organic matter by aerobic microorganisms, known as respiration, 239 confirmed microbial activity in all the soils. According to [21], soil respiratory activities and microbial 240 abundance are sensitive to contamination with petroleum derivatives. [22] Associated decline of 241 respiratory activity similar to what is represented in Table 5 to depleted available carbon substrates. 242 Additionally, respiration declines in soils that lack nutrients and other supporting factors for microbial and other biological activities [23]. While respiratory activity showed a substantial negative correlation 243 244 with OM for the unpolluted control (-0.96 PCC), its correlation with OM for polluted control showed 245 arguably no correlation (-0.07 PCC). Nonetheless, a fair positive correlation was recorded in S. 246 americana treated and S. ocymoides treated soils (+0.51 and +0.54 PCC, respectively).

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248Table 5:Respiratory activity (CO2.C) (in mg) of unpolluted control, polluted control,249Schwenkia americana and Spermacoce ocymoides.

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Group	Before	Week 12
Unpolluted	1.4400±0.01 ^a	0.800±0.37 ^{a*}
Control		
Polluted	1.9200±0.01 ^b	0.24±0.12 ^{a*}
Control		
Schwenkia	1.9200±0.01 ^b	0.40±0.37 ^{b*}
americana		
Spermacoce	1.9200±0.01 ^b	0.28±0.18 ^{c*}
ocvmoides		

251 Values are mean ± standard deviations of triplicate determinations.

Values in the same column with different letters (a,b) are significantly different at P = .05.

253 *P = .05 compared to the corresponding values before treatment.

254 255 As shown in Table 6, the soil organic matter in the remediated and polluted control groups, when 256 compared with the baseline values, reduced over time. Organic matter is the major source of plant 257 nutrients [24] and its mineralization depends on the interaction between the chemicals present in the 258 soil [4]. [25] Reported that the decomposition of organic matter is largely a biological process that 259 occurs naturally and determined by soil organisms, the physical environment and the quality of the organic matter. The reduction in organic matter in the groups may therefore be associated with its 260 261 utilization by the microorganisms to release nutrients for use by plants and microorganisms. 262

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 Table 6:
 Organic matter (in %) of unpolluted control, polluted control, Schwenkia

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 americana and Spermacoce ocymoides.

Group	Before	Week 8	Week 12
Unpolluted Control	2.4900±0.01 ^a	2.4700±0.22 ^a	2.2267±0.19 ^a
Polluted	4.800±0.10 ^b	4.0200±0.09 ^{b*}	3.7767±0.14 ^{b*}

Control			
Schwenkia	4.800±0.10 ^b	3.8467±0.24 ^{b*}	3.7267±0.11 ^{b*}
americana Spermacoce	4.800±0.10 ^b	3.8067±0.25 ^{b*}	3.5767±031 ^{b*}
, ocymoides			

266 Values are mean ± standard deviations of triplicate determinations.

267 Values in the same column with different letters (a,b) are significantly different at P = .05.

268 **P* = .05 compared to the corresponding values before treatment.

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 Table 7:
 Pearson's correlation coefficient (PCC) of observed enzyme activities versus

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 organic matter (OM)

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Unpolluted control	Polluted control	S. americana	S. ocymoides
-1.00*	-0.98	-0.80	+0.47
+0.36	-0.90	+0.96	+0.55
+0.84	+0.05	+0.80	+0.55
+0.61	+0.52	+0.36	-0.57
-0.96	-0.07	+0.51	+0.54
	Unpolluted control -1.00* +0.36 +0.84 +0.61 -0.96	Unpolluted control Polluted control -1.00* -0.98 +0.36 -0.90 +0.84 +0.05 +0.61 +0.52 -0.96 -0.07	Unpolluted control Polluted control S. americana -1.00* -0.98 -0.80 +0.36 -0.90 +0.96 +0.84 +0.05 +0.80 +0.61 +0.52 +0.36 -0.96 -0.07 +0.51

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275 4. CONCLUSION

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277 Crude oil spillage presents deleterious effects on the environment. Both microbial activities of the soil 278 can reflect sensitively the quality of soil, and soil enzyme activities can directly reflect the metabolic 279 need and nutrient availability of soil microorganisms which are important key nutrients' processing and 280 recovery from detrital inputs and accumulated soil organic matter. Microorganisms secrete 281 degradative enzymes which can counter the effect poised by the spillage thus effecting amelioration 282 of the pollutants' effects in the polluted soil. The extracellular enzymes; protease, dehydrogenase, 283 acid and alkaline phosphatase activities are shown to vary with crude oil pollution relative to time thus 284 indicating ameliorative effects.

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